



# IJSRM

INTERNATIONAL JOURNAL OF SCIENCE AND RESEARCH METHODOLOGY

An Official Publication of Human Journals



Human Journals

Research Article

June 2021 Vol.:18, Issue:4

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## Adsorption of Chromium (VI) and Orange Methyl onto Activated Carbon Obtained from The Cores of *Canarium ovatum*: Influence of Functionalization



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**Submitted:** 22 May 2021

**Accepted:** 28 May 2021

**Published:** 30 June 2021



[www.ijsrm.humanjournals.com](http://www.ijsrm.humanjournals.com)

**Keywords:** Functionalization, Simultaneous Adsorption, Boehm Analysis, Pollutant, Affinity

### ABSTRACT

This work highlights the effect of the functionalization of activated carbon with sulfuric acid on its affinity towards inorganic and organic pollutants during simultaneous adsorption. The virgin activated carbon was obtained by chemical activation with phosphoric acid, then the modified activated carbon was obtained by carrying out a 20% functionalization with sulfuric acid. The lignocellulosic nature of the plant material used was verified by thermogravimetric analysis. The two adsorbents were characterized by Fourier transform infrared analysis; evaluation of functional groups of surfaces using Boehm analysis and porosity parameters namely iodine number and methylene blue number. The isotherms of Langmuir, Freundlich, Temkin, DR and Frumkin were studied. The kinetics were studied using pseudo-first order, pseudo-second order, Elovich, and intraparticle models. Fourier transform infrared analysis and Boehm analysis show an increase in phenolic groups; carboxylic and lactonic acids on the surface of the modified activated carbon. The results obtained reflect a drop in the specific surface area through the very significant drop in the iodine number (from 1422.40mg/g to 381.00mg/g). The Langmuir isotherm has the highest values of the correlation coefficient at the surface of the modified activated carbon (0.9943 and 0.9563), the Freundlich isotherm has the highest  $R^2$  for the adsorption of Cr (VI) at the surface of activated carbon (0.9037) and the Temkin isotherm has the highest  $R^2$  for adsorption of orange methyl to the surface of activated carbon (0.9869). The pseudo-first order model best describes the chromium (VI) adsorption kinetics at the surface of virgin activated carbon with the highest  $R^2$  (0.6947). While the latter is better described on the surface of activated carbon modified by the pseudo-second order model ( $R^2 = 0.9944$ ). Elovich model has the highest correlation coefficient values for the adsorption of methyl orange to the surface of virgin activated carbon and modified activated carbon (0.9929 and 0.9905, respectively).

## INTRODUCTION

The release into the water of pollutants with various natures and toxicities remains a major concern. This precious liquid nevertheless represents an incomparable stake in the life of living beings. Among the most dangerous contaminants that can be found in water, we have metal ions such as Cr (VI) and organic dyes such as orange methyl. These can be found in nature as a result of rock erosion and soil leaching for Cr (VI) or due to industrial activities [1]. These pollutants have mutagenic and carcinogenic properties and are highly toxic to living beings [2]. The World Health Organization also sets the limit for the concentration of Cr (VI) in water at 0.05mg/l and that of orange methyl at 3.1mg/l [3, 4]. One of the methods of treating water loaded with Cr (VI) is reduction in Cr (III), however, this requires expensive reagents. On the other hand, this reduction is incomplete and thus leads to obtaining toxic sludge [5, 6]. The adsorption on activated carbon is therefore presented as the best method of water decontamination because it has the advantage of being very effective on a wide range of pollutants [7]. The studies currently carried out relate to the various techniques for improving the properties of activated carbon. Several studies have therefore been carried out on the modification of the surface of activated carbon with noble metals, namely silver; palladium; platinum, or with essential heavy metals including iron, as well as modification by oxidation with agents such as sulfuric acid and nitric acid [7-9]. These treatments can be at the origin of the creation or the multiplication of oxygen functions which could have an impact on the porous properties on the surface of the adsorbent material [10, 11]. Indeed, certain works mentioned in the literature show that an acid functionalization of activated carbon leads to the reduction of the specific surface area and the development of surface chemical functions, such as -OH; -COOH; -CO and -SO<sub>3</sub>H functions [12-14]. All these parameters have an influence on the adsorption behavior of activated carbon [12, 15, 16]. In this wake, studies have been carried out on the functionalization of carbon nanotubes with nitric acid for the adsorption of Cr (VI) [6]. These help to say that such a treatment promotes an ion exchange phenomenon between the -OH found on the surface of the activated carbon and the various anionic forms of Cr (VI) in solution [17-19].

The objective of this work is to synthesize an activated carbon functionalized with sulfuric acid and to study its affinity towards organic pollutants such as orange methyl and metal ions such as Cr(VI). Functionalized activated carbon will be characterized by infrared analysis; the Boehm

titration; the evaluation of the pH of the zero charge point as well as the porosity indicators namely the methylene blue number and the iodine number. The effect of functionalization will be evaluated for the adsorption of orange methyl and Cr (VI).

## **MATERIALS AND METHODS**

Phosphoric acid 85% ( $\text{H}_3\text{PO}_4$ ), sodium hydroxide (NaOH), hydrochloric acid 36% (HCl) and sulfuric acid 98% ( $\text{H}_2\text{SO}_4$ ) were supplied by JHD Laboratories; sodium bicarbonate ( $\text{NaHCO}_3$ ) and sodium carbonate ( $\text{Na}_2\text{CO}_3$ ) are produced by BIOCHEM laboratories; sodium thiosulphate ( $\text{Na}_2\text{S}_2\text{O}_3$ ) and sodium chloride (NaCl) are supplied by KERMEL laboratories; diiodine ( $\text{I}_2$ ) and potassium iodide (KI) are obtained from RIEDEL-DE HAEN laboratories; methylene blue and methyl orange are supplied by TRUST CHEMICAL ECO laboratories; potassium dichromate ( $\text{K}_2\text{Cr}_2\text{O}_7$ ) and starch are produced by JHD Laboratories.

### **Preparation of activated carbon**

The activated carbon used is obtained from the cores of *Canarium Ovatum* collected in the locality of Mbouda (10°12" east longitude and 5°30" north latitude) in the region of western Cameroon. This is synthesized by chemical activation with phosphoric acid.

The cores of *Canarium Ovatum* are washed thoroughly with distilled water and then dried at room temperature before being crushed and sieved to obtain a powder with a particle size of less than 250 $\mu\text{m}$ . The powder obtained was impregnated with phosphoric acid according to the powder/activator mass ratio 1/3, dried at 105°C. The resulting solid is calcined at 450°C in a NABERTHERN brand electric furnace with a temperature rise of 5°C/min and a residence time of one hour. After this calcination, the carbon is allowed to cool to room temperature in the oven before removal. The activated carbon obtained is washed with distilled water until obtaining water with a pH close to 7, it is finally dried in an oven at 105°C. After this procedure we obtain activated carbon (AC).

### **Functionalization of the activated carbon obtained**

20% functionalization using sulfuric acid was performed. A mass of 25g of activated carbon is introduced into a volume of 100ml of a sulfuric acid solution of concentration 0.5g/l. The whole is stirred for 4 hours before being dried at 105°C. The solid resulting from this step is washed

with hot distilled water until washing water of neutral pH is obtained. The bio-adsorbent obtained is the activated carbon modified by sulfuric acid (ACMSA).

### Characterization of bio-adsorbents

Fourier transform infrared analysis was carried out to determine the different chemical groups present on the surface of our materials.

The Boehm titration was performed to evaluate the different acidic and basic chemical functions on the surface of each of the activated carbons studied. Thus, a mass of 0.1g of each adsorbent was introduced into a volume of 50ml of respective solutions of NaOH, NaHCO<sub>3</sub>, and Na<sub>2</sub>CO<sub>3</sub> with a concentration of 0.1N. The mixtures are stirred for 72 hours and then filtered. A volume of 5 ml of the filtrate is collected and assayed with a 0.1N solution of hydrochloric acid in the presence of helianthine. NaOH identifies carboxylic acids; lactonic and phenolic, Na<sub>2</sub>CO<sub>3</sub> is used to determine lactonic and carboxylic acids while NaHCO<sub>3</sub> only corresponds to carboxylic acids. The basic functions are determined globally using a volume of 50ml of a 0.1N solution of hydrochloric acid, following the same procedure. The titration is carried out this time with a solution of NaOH of concentration 0.1N in the presence of phenolphthalein. The formula for quantifying each chemical function is as follows [20]:

$$N_0V_0 - N_fV_0 = neqgR \quad (1)$$

$N_0V_0$ : Number of gram equivalents before the reaction,  $N_fV_0$ : Number of gram equivalent after the reaction,  $neqgR$ : Number of gram equivalent reacted.

The charge on the surface of the materials was evaluated by determining the pH of the zero charge point (pH<sub>zcp</sub>). This parameter is determined by introducing 0.1g of activated carbon into a volume of 50 ml of sodium chloride solutions with a concentration of 0.1N and a pH varying from 2 to 10. The whole is stirred for 72 hours. The pH of the solutions is adjusted using solutions of NaOH (0.1N) and HCl (0.1N). The pH<sub>zcp</sub> is obtained by identifying the point of intersection between the bisector and the curve  $pH_f = f(pH_i)$  [20].

The mesoporous capacity of the adsorbents was estimated by determining the methylene blue number. A mass of 0.3g of adsorbent was introduced into a volume of 100ml of a methylene blue solution with a mass concentration of 1000mg/l. The whole is stirred for 2 hours and then

filtered; the filtrate is passed through a spectrophotometer at a wavelength of 660nm. The methylene blue number is given by the formula:

$$I_{BM}(\text{mg/g}) = \frac{(C_i - C_f)}{m_{CA}} \times 0.1 \quad (2)$$

The mesoporous surface can be approximated by applying the formula [21]:

$$S_{BM} = \frac{Q^{\infty} S N_A}{M_{BM}} \quad (3)$$

$C_i$ : initial concentration of the solution,  $C_f$ : residual concentration of the solution,  $m_{CA}$ : mass of activated carbon,  $S_{BM}$ : area of methylene blue in  $\text{m}^2/\text{g}$ ,  $Q^{\infty}$ : maximum quantity in  $\text{mg/g}$  of adsorbed methylene blue obtained by the Langmuir isotherm,  $S$ : area occupied by a methylene blue molecule ( $S = 175\text{\AA}^2$ ),  $N_A$ : Avogadro number ( $N_A = 6.02 \times 10^{23} \text{mol}^{-1}$ ),  $M_{BM}$ : molar mass of methylene blue ( $M_{BM} = 319.859 \text{g/mol}$ ).

The estimation of the microporous capacity is made by determining the iodine number. A mass of 0.1g of adsorbent is introduced into a volume of 20ml of iodine solution ( $N = 0.02\text{N}$ ). The whole is stirred for 3 hours and then filtered and 10ml of the filtrate is taken before being determined using a 0.1N sodium thiosulfate solution. 2 drops of starch are used as an indicator of the change from the blue color of the solution to colorless marking the end of the dosage. A blank test should be performed under the same conditions without activated carbon. The iodine number is given by the relation [22]:

$$I_2(\text{mg/g}) = \frac{V_b - V_s}{m_{CA}} \times 12,7 \times 2 \quad (4)$$

$V_b$ : volume of thiosulfate for the blank,  $V_s$ : volume of thiosulfate for the adsorbed solution,  $m_{CA}$ : mass of activated carbon.

### Batch adsorption

The different solutions of chromium (VI) and orange methyl will be determined using a UV-visible spectrophotometer at the respective wavelengths of 354nm and 460nm. The adsorbed quantity will be obtained thanks to the relation:

$$Q_e(\text{mg/g}) = \frac{(C_i - c_e)}{m_{CA}} \times V_{ads} \quad (5)$$

$Q_e$ : amount adsorbed at equilibrium,  $C_i$ : initial concentration in mg/l,  $C_e$ : residual concentration in mg/l,  $V_{ads}$ : volume adsorbed in ml,  $m_{CA}$ : mass of activated carbon in g.

The study of the effect of some parameters such as contact time (5min to 65min); the mass of adsorbent (0.03g to 1.2g); the pH of the solution (2 to 10) and the initial concentration (100mg/l to 300mg/l) was carried out at room temperature.

### Isotherms study

The adsorption equilibrium of our various bio-adsorbents has been studied using isotherms from Langmuir, Freundlich, Temkin, Dubinin-Radushkevich and Frumkin.

Langmuir isotherm [23]:

The linear form of this isotherm is:  $\frac{1}{q_e} = \frac{1}{Q_{max}} + \frac{1}{Q_{max} b C_e}$  (6)

$$R_L = \frac{1}{1 + b C_0} \quad (7)$$

$q_e$ : quantity adsorbed at equilibrium in mg/g,  $Q_{max}$ : maximum quantity in mg/g of pollutant adsorbed in a monolayer at the surface of the adsorbent,  $b$ : Langmuir constant in l/mg linked to free energy,  $C_e$ : residual concentration at equilibrium in mg/l,  $R_L$ : Langmuir constant ( $R_L > 1$  unfavorable adsorption,  $R_L = 1$  linear adsorption,  $0 < R_L < 1$  favorable adsorption,  $R_L = 0$  irreversible adsorption),  $C_0$ : concentration initial pollutant expressed in mg/l.  $Q_{max}$  and  $b$  are deduced from the intercept and the slope on the plot of  $\frac{1}{q_e}$  as a function of  $\frac{1}{C_e}$ .

Freundlich isotherm [23]:

The linearization of its equation is given by the formula:

$$\log q_e = \log K_F + \frac{1}{n} \log C_e \quad (8)$$

$K_F$ : Freundlich constant indicating the adsorption capacity,  $1/n$ : Freundlich constant indicating the adsorption intensity ( $1/n > 1$  unfavorable adsorption,  $1/n = 1$  homogeneous adsorption and no

interaction between species,  $0 < 1/n < 1$  favorable adsorption); it corresponds to the slope on the plot of the line of  $\log q_e$  as a function of  $\log C_e$ .

Temkin isotherm [24]:

The linearization of the basic equation leads to the formula:

$$q_e = \frac{RT}{b_T} \ln A + \frac{RT}{b_T} \ln C_e \quad (9)$$

R: ideal gas constant ( $R=8,314\text{J}\cdot\text{mol}^{-1}\cdot\text{K}^{-1}$ ), T: absolute temperature in °K,  $b_T$ : heat of adsorption in  $\text{J}\cdot\text{mol}^{-1}$ , A: constant corresponding to the energy of maximum equilibrium binding in  $\text{l}\cdot\text{mg}^{-1}$ . A and  $b_T$  are deduced from the slope and the intercept on the plot of the line representing  $q_e$  as a function of  $\ln C_e$ .

Dubin Radushkevich Isotherm [25]:

The linear form is given by the equation:

$$\ln q_e = \ln q_m - K\varepsilon^2 \quad (10)$$

Where  $\varepsilon = RT \ln(1 + \frac{1}{C_e})$  (11)



$$E_a = -\frac{2}{3} \sqrt{2K} \quad (12)$$

K: constant related to the activation energy in  $\text{mol}^2\cdot\text{KJ}^{-2}$  corresponding to the slope on the plot representing  $\ln q_e$  as a function of  $\varepsilon^2$ , R: ideal gas constant in  $\text{KJ}\cdot\text{mol}^{-1}\cdot\text{K}^{-1}$ ,  $E_a$ : activation energy in  $\text{KJ}\cdot\text{mol}^{-1}$ .

Frumkin isotherm [24]:

The linear equation for this isotherm is as follows:

$$\ln \left[ \left( \frac{\theta}{1-\theta} \right) \frac{1}{C_e} \right] = \ln K + 2a\theta \quad (13)$$

With:  $\theta = \frac{q_e}{q_m}$  (14)



$$\ln K = \frac{-\Delta G}{RT} \quad (15)$$

K: constant related to the adsorption equilibrium,  $\Delta G$ : free energy of adsorption in  $\text{KJ.mol}^{-1}$ ,  $q_m$ : theoretical maximum capacity of adsorption in monolayer provided by the isotherm DR in  $\text{mg/g}$ ,  $a$ : interaction coefficient between adsorbed particles,  $a > 0$ : attraction between adsorbed particles;  $a < 0$ : repulsion between adsorbed particles;  $a = 0$  no interaction between adsorbed particles). These different constants are obtained from the plot of  $\ln \left[ \left( \frac{\theta}{1-\theta} \right) \frac{1}{C_e} \right]$  as a function of  $\theta$ .

### Study of adsorption kinetics

The kinetic models studied in this work are the pseudo-first order model; the pseudo-second order model; the intra-particle diffusion model and the Elovich model.

Pseudo-first-order model [24]:

This model is governed by the relation:

$$\ln(q_e - q_t) = \ln q_e - K_1 t \quad (16)$$

$q_e$  and  $q_t$  are the amounts in  $\text{mg.g}^{-1}$  adsorbed at equilibrium and at time  $t$ ,  $K_1$  is the adsorption rate constant in  $\text{ml.min}^{-1}$ .  $q_e$  and  $K_1$  are obtained from the plot of the line representing  $\ln(q_e - q_t)$  as a function of  $t$ .

Pseudo-second-order model [26]:

Its equation is as follows:

$$\frac{1}{q_t} = \frac{1}{K_2 q_e^2} + \frac{t}{q_e} \quad (17)$$

$K_2$  corresponds to the rate constant in  $\text{g.m}^{-1}.\text{min}^{-1}$  obtained from the intercept on the plot of  $\frac{1}{q_t}$  as a function of  $t$ .

The initial rate  $h$  in  $\text{mg.g}^{-1}.\text{min}^{-1}$  is given by the relation:  $h = K_2 q_e^2$  (18)

Intra-particle model [27]:



Its equation is as follows:

$$q_t = K_p t^{1/2} + C \quad (19)$$

$K_p$  is the intra-particle diffusion constant expressed in  $\text{mg.g}^{-1}.\text{min}^{-1/2}$ ,  $C$  is the thickness of the diffusion layer. They are obtained by identifying the slope and the intercept on the plot representing  $q_t$  as a function of  $t_{1/2}$ .

Elovich model [28]:

This model is governed by equation:

$$q_t = \frac{1}{\beta} \ln \alpha \beta + \frac{1}{\beta} \ln t \quad (20)$$

$\alpha$  is the initial adsorption rate constant in  $\text{mg.g}^{-1}.\text{min}^{-1}$  and  $\beta$  is the desorption constant in  $\text{g.mg}^{-1}$ . These constants are obtained by plotting the line  $q_t$  as a function of  $\ln t$ .

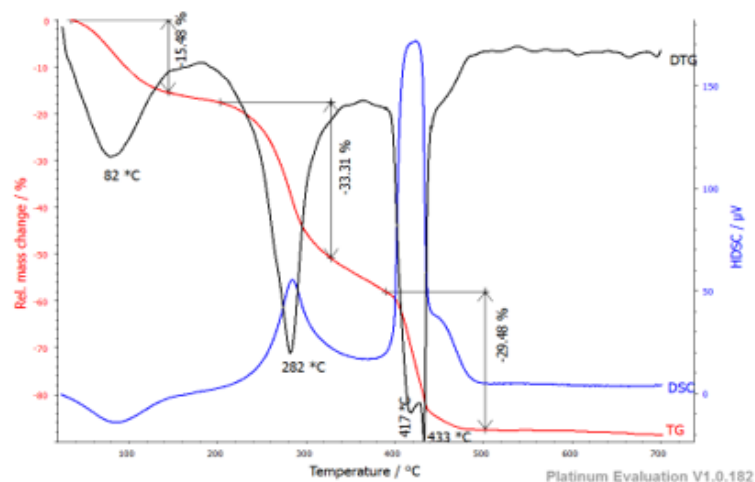
## RESULTS AND DISCUSSION

### Characterization of the adsorbents

#### Thermogravimetric analysis

This analysis was carried out in the laboratories of the Faculty of Sciences of the University of Yaoundé. The thermogram in figure no 1 shows the gradual decomposition of the plant precursor as a function of the temperature.



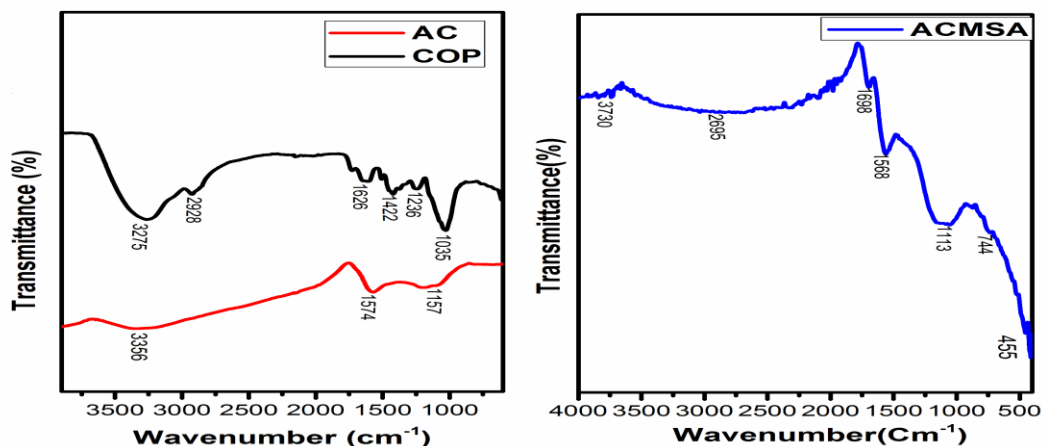


**Figure no. 1: TGA / DSC of Canarium Ovatum powder**

This decomposition takes place in three stages: a mass loss of 15.48% at 82°C which could correspond to the disappearance of the physisorbed moisture on the plant material surface [29]; the decomposition of hemicellulose and cellulose at 282°C resulting in a mass loss of 33.31% and finally, the mass loss of 29.48% observed at 417°C could correspond to thermal decomposition lignin. From 450°C, an almost constant evolution of the mass is observed; there is no more loss of mass. This confirms the lignocellulosic nature of the plant material and makes it possible to set the calcination temperature for obtaining the activated carbon at 450°C.

### **Fourier transform infrared analysis (FTIR)**

Figure no 2 shows the infrared spectra of Canarium Ovatum powder (COP) as well as virgin activated carbon (AC) and Activated Carbon Modified with Sulfuric acid (ACMSA).



**Figure no 2: Fourier transform infrared analysis of COP, AC and ACMSA**

Several peaks present on the FTIR spectrum of COP disappear on that of AC: this is the case of the peak present at  $1422\text{cm}^{-1}$  and the one present at  $1236\text{cm}^{-1}$ . While on the FTIR spectrum of ACMSA we observe an increase in the intensity of peaks already present on the FTIR spectra of AC followed by the appearance of new peaks. Thus, the broad peak present at  $3500\text{cm}^{-1}$  on the FTIR spectrum of COP is only visible at very low intensity at  $3356\text{cm}^{-1}$  on the FTIR spectrum of AC. This peak corresponds to the OH group of carboxylic acids, alcohols, or water molecules absorbed by plant material [30]. The sharp drop in intensity observed on the FTIR peak of AC would be due to the loss of part of the OH group in the form of an evaporated water molecule during the process of calcination of the plant material as predicted by the thermogravimetric analysis. We also observe a peak at  $2928\text{cm}^{-1}$  which reflects the aliphatic C-H. The peaks present at  $1626\text{cm}^{-1}$  and  $1035\text{cm}^{-1}$  on the FTIR spectrum of COP appear instead at  $1574\text{cm}^{-1}$  and  $1113\text{cm}^{-1}$  on the FTIR spectrum of AC with a sharp decrease in intensity. These peaks relate to the C = O group of carboxylic acids, ketones, and aldehydes; the decrease in the intensity of the latter would be due to the loss of part of these groups in the form of carbon dioxide during calcination. The peaks present at  $1422\text{cm}^{-1}$  and  $1236\text{cm}^{-1}$  on the FTIR spectrum of the plant material are respectively the peaks of deformation in the plane of acid OH and of the C-O-C groups of ester and ethers functions. Observation of the FTIR spectrum of modified activated carbon shows us a large, intense peak between  $2400\text{cm}^{-1}$  and  $3730\text{cm}^{-1}$  characteristic of the OH group of carboxylic acids and phenols [31]. The peak visible at  $1698\text{cm}^{-1}$  corresponds to the C = O of the lactones, the literature places this peak between  $1620\text{cm}^{-1}$  and  $2110\text{cm}^{-1}$  [32]. The peaks

present respectively at  $1568\text{cm}^{-1}$  and  $1113\text{cm}^{-1}$  were already visible on the FTIR spectrum of the unmodified activated carbon at  $1574\text{cm}^{-1}$  and  $1157\text{cm}^{-1}$ , we just notice an increase in the intensity of these peaks following modification with sulfuric acid. This behavior is moreover visible on all the peaks of the FTIR spectra of ACMSA and is explained by the oxidizing character of sulfuric acid which tends to promote the development of oxygen functions which is manifested in the FTIR spectra by increasing peak intensities [33]. The literature associates the peak at  $744\text{cm}^{-1}$  with out-of-plane distortion vibrations of aromatic C-H groups [34].

**Physico-chemical characteristics of activated carbon:**

The physicochemical characteristics of AC and ACMSA are presented in Table no 1.

**Table no. 1: physicochemical characteristics of AC and ACMSA**

	AC	ACMSA
<b>Id<sub>BM</sub> (mg/g)</b>	318.99	327.35
<b>S<sub>BM</sub> (m<sup>2</sup>/g)</b>	499.04	569.63
<b>Id<sub>I2</sub> (mg/g)</b>	1422.40	381.00
<b>pH<sub>ZCP</sub></b>	6.00	5.80

We observe from Table 1 a low increase in the methylene blue number as well as in the mesoporous surface area and a very significant decrease in the iodine number which must necessarily lead to a sharp decrease in the surface area occupied by the micropores. This decrease in porosity could be explained by the fact that the oxygenated functions developed form on the surface of the pores, thus leading to the narrowing of the latter [33]. On the other hand, we also observe a decrease in the pH of the zero charge point of the activated carbon functionalized with sulfuric acid compared to that of virgin activated carbon this would mean that the modified activated carbon is more acidic than the unmodified activated carbon. . This result is verified by the result provided by the infrared analysis which shows an increase in oxygen functions on the surface of ACMSA responsible for the acidity.

**Boehm titration**

Table no 2 summarizes the results of the Boehm titration of virgin activated carbon and modified activated carbon.

**Table no 2: Surface functions of AC and ACMSA**

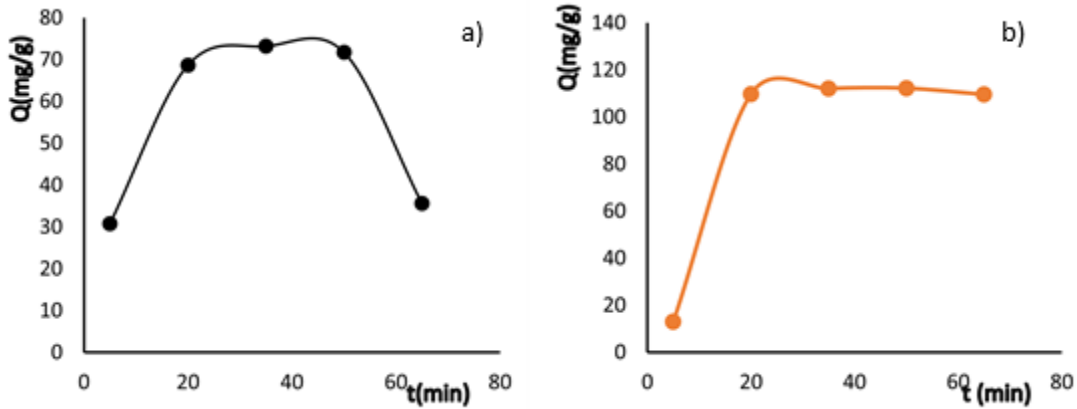
		Quantity of Functional group (meqg/g)	Total group (meqg/g)
<b>AC</b>	Carboxylic acid	0.01	0.028
	Phenolic acid	0.013	
	Lactonic Acid	0.005	
	Base	0.009	0.009
<b>ACMSA</b>	Carboxylic acid	0.016	0.034
	Phenolic acid	0.012	
	Lactonic Acid	0,006	
	Base	0,002	0.002

We observe from this table no 2 an increase in the amount of acidic functional groups when switching from virgin activated carbon to modified activated carbon (from 0.028meqg/g to 0.034meqg/g) and a large decrease in the quantity of functional groups basic (from 0.009meqg/g to 0.002meqg/g). The increase in acidity at the surface of CAAS is mainly due to the increase in the quantity of functional groups of carboxylic acids and phenolic acids. Sulfuric acid is therefore believed to be responsible for the oxidation of certain functional groups such as the quinones responsible for the basicity on the surface of the activated carbon [33]. This result justifies the pH<sub>zcp</sub> values found.

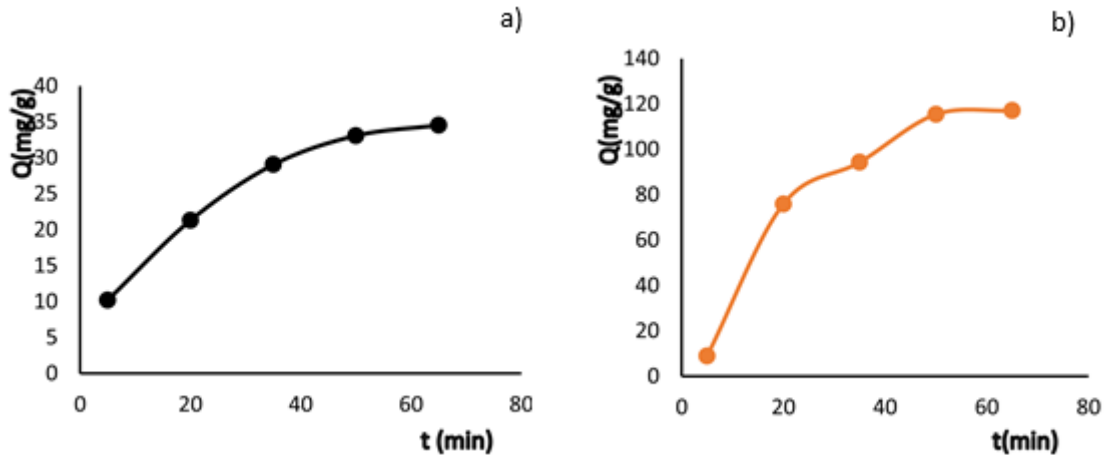
### **Adsorption tests**

#### **Contact time effect**

The effect of time on the simultaneous removal of chromium (VI) and orange methyl is shown in figure no 3 and figure no 4.



**Figure no. 3: Effect of contact time on the adsorption capacity of Cr (VI) (a) and orange methyl (b) onto AC. Pollutants concentration = 100mg/l, pH = 2.**



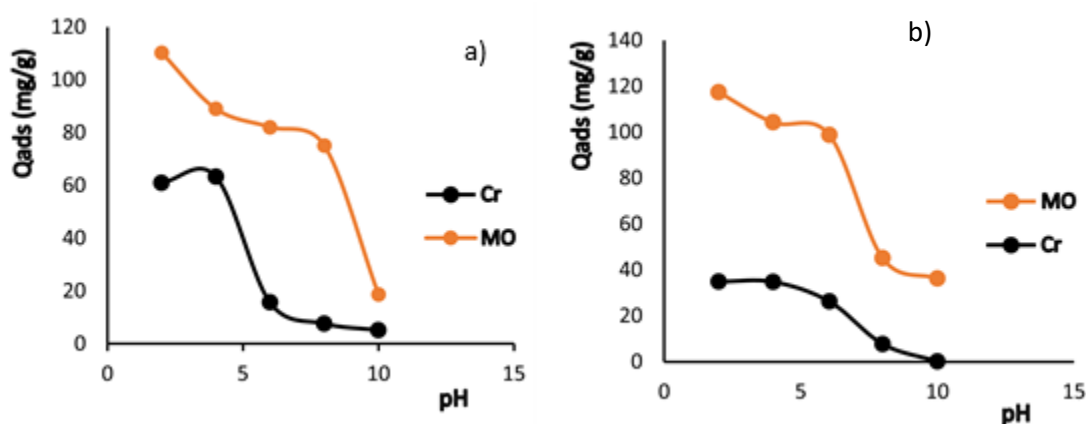
**Figure no 4: Effect of contact time on adsorption capacity of Cr (VI) (a) and orange methyl (b) onto ACMSA. Pollutants concentration = 100mg/l, pH = 2.**

We generally observe a rapid increase in the pollutant adsorption capacity onto our two adsorbents during the first twenty minutes; this could be explained by the full availability of adsorption sites at the start of the process. Beyond this time we observe a relative slow adsorption which may be due to the migration of pollutant particles to internal adsorption sites resulting from the occupation of external sites. For AC (figure no 3), this behavior can be observed until the thirty-fifth minute when a tray is formed, reflecting the saturation of the sites on the surface of the adsorbent: the adsorption equilibrium is reached; this saturation also leads to desorption of Cr (VI) (figure no 3a). Better behavior is observed with ACMSA (figure no 4)

where saturation occurs from the sixtieth minute and no desorption is visible. This large action time for ACMSA could be explained by the development of more oxygenated functions on the surface of the latter which would intervene in the adsorption process following a phenomenon of ion exchange between the OH of the acid functions and the anionic form of chromium [34, 35].

### Effect of solution pH

Figure no 5 shows the effect of pH of the solution on the adsorption capacity of Cr (VI) and orange methyl.



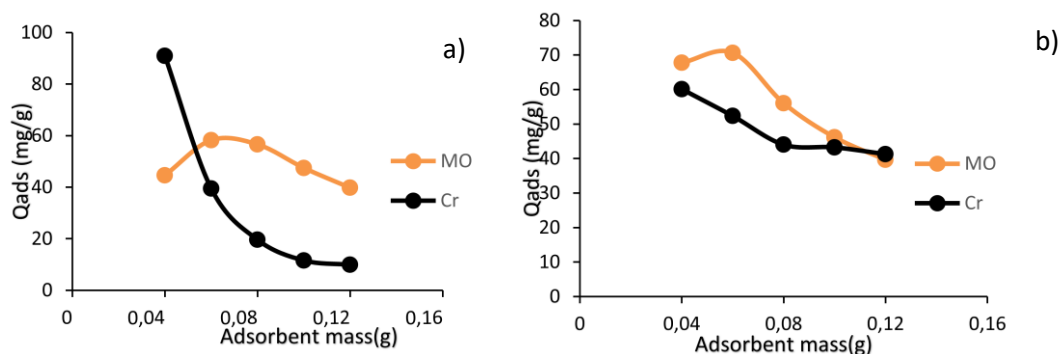
**Figure no. 5: Effect of pH on the adsorption capacity of Cr(VI) and orange methyl onto AC (a) and ACMSA (b). Pollutants concentration = 100mg/l, Contact time = 35min.**

The initial pH of the solution influences the charge on the surface of the adsorbent material and the ionic form of the pollutant [36]. The adsorption to the surfaces of AC and ACMSA is maximum at pH 2, it decreases with increasing pH and becomes very low in strongly basic media (pH>8). This could be explained by the fact that at low pH values our materials have a positively charged surface; Cr(VI) is present as an anion in aqueous medium ( $\text{HCrO}_4^-$  and  $\text{CrO}_4^{2-}$ ) as well as orange methyl [37], adsorption could be easily favored by the establishment of electrostatic strengths between the particles of the different forms of Cr(VI), orange methyl and the respective surfaces of the activated carbon. In aqueous medium with a basic pH, the presence of excess  $\text{OH}^-$  ions creates competition between those  $\text{OH}^-$  and anionic pollutants present in the medium, which would explain the drop in adsorption capacity at the respective surfaces of AC and ACMSA.



### Effect of adsorbent mass

Figure no. 6 shows the effect of adsorbent mass on the adsorption capacity of Cr(VI) and orange methyl onto AC and ACMSA.

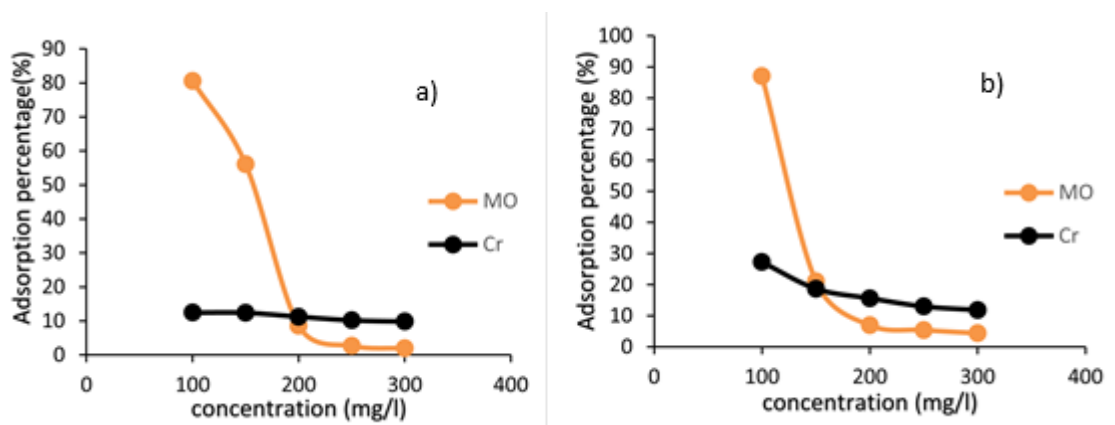


**Figure no. 6: Effect of adsorbent mass on the adsorption capacity of Cr (VI) and orange methyl onto AC (a) and ACMSA(b). Pollutants concentration = 100mg/l, Contact time = 35min, pH = 2.**

The absorption capacity of Cr(VI) and orange methyl onto AC and ACMSA decreases. This could be explained by the fact that the increase in the mass of adsorbent leads to the aggregation of adsorption sites making them inaccessible [38]. However, we find that if the adsorption of Cr(VI) is maximum with a mass of activated carbon equal to 0.04g, that of orange methyl is maximum with a mass of activated carbon equal to 0.05g.

### Effect of initial concentration:

Figure no. 7 shows the effect of the initial pollutant concentration on the removal percentage at the surface of our respective adsorbents.



**Figure no. 7: Effect of the initial pollutant concentration on the removal percentage of Cr(VI) and orange methyl onto AC (a) and ACMSA (b). Contact time = 5min, pH = 2.**

We see a decrease in the absorption percentage of Cr(VI) and orange methyl when we move from low concentrations (100mg/l) to high initial concentrations (300mg/l) regardless of the activated carbon. Indeed the adsorption percentage goes from 80% and 90% to almost zero and 5% respectively for AC and ACMSA about the adsorption of orange methyl and from 10% and 30% to 8% and 10% respectively for AC and ACMSA concerning the adsorption of Cr(VI). This could be explained by the fact that we have an increase in the pollutant concentration for a constant amount of activated carbon and therefore for a constant specific surface [39]. Furthermore, we note that the modification of the surface of the activated carbon with sulfuric acid is at the origin of the increase in the adsorption percentage of 10% for orange methyl and 20% for Cr(VI).

### Adsorption isotherms

The mathematical description of the adsorption equilibrium gives useful information on the adsorption mechanism, the surface properties of the adsorbent as well as the affinity between the adsorbate and the adsorbent thanks to the parameters obtained from the straight lines regressions [40]. These parameters are grouped in Table no 3.

Table no. 3: Isotherm parameters of Cr(VI) and orange methyl adsorption onto ACMSA and AC

	ACMSA		AC	
	Cr (VI)	MO	Cr (VI)	MO
<b>Langmuir</b>				
$Q_{maxexp}(mg/g)$	36.9000	117.4347	44.6916	112.2336
$Q_e (mg/g)$	107.5268	196.0784	46.5116	18.4501
$b (L/mg)$	0.0019	0.0078	0.0340	0.0900
$R_L$	0.8349	0.5597	0.1246	
$R^2$	0.9943	0.9562	0.6699	
<b>Freundlich</b>				
$K_f$	$3.6975 \cdot 10^{-7}$	0.1950	3.7235	0.0014
$1/n$	0.7407	0.5380	0.2033	0.6337
$R^2$	0.9847	0.9447	0.9037	0.9412
<b>Temkin</b>				
$b_T (J/mol)$	316.3719	55.6004	135.8839	-79.0193
$A (l/mg)$	0.9050	0.0659	0.0268	0.0024
$B_T$	7.8312	46.2230	18.2330	-31.3540
$R^2$	0.9912	0.9532	0.8877	0.9869
<b>DR</b>				
$E (KJ/mol)$	21.3200	35.3553	0.0497	
$Q_e (mg/g)$	36.9400	117.4367	41.6916	20.6910
$K (mol^2/KJ^2)$	0.0011	0.0004	202.2400	-49.0760
$R^2$	0.9630	0.9275	0.6550	0.8093
<b>Frumkin</b>				
$K$	$9.2444 \cdot 10^{-5}$	$4.8648 \cdot 10^{-5}$	$3.3552 \cdot 10^{-6}$	0.0172
$a$	4.1670	4.3693	4.3900	4.2304
$\Delta G (KJ/mol)$	23.0192	24.6045	31.2297	10.0650
$R^2$	0.6300	0.9061	0.8733	0.6125

In view of the results obtained, the Langmuir isotherm has the highest  $R^2$  for the adsorption of orange methyl to the surface of ACMSA (0.956). This also has a correlation coefficient close to 1 for the adsorption of Cr(VI) to the surface of that adsorbent (0.9943). The Freundlich isotherm has the highest  $R^2$  for adsorption of Cr(VI) to the surface of virgin activated carbon (0.9037). Temkin isotherm has the highest correlation coefficient value for the adsorption of orange methyl to the surface of AC (0.9869). In view of this, we deduce that the Langmuir isotherm best describes the adsorption of the two pollutants to the surface of the modified activated carbon while the Freundlich and Temkin isotherms are those which best describe the adsorption of Cr(VI) and orange methyl respectively on the surface of virgin activated carbon. On the other hand, the values between zero and one of Langmuir's constant  $R_L$  indicate favorable adsorption of the two pollutants to the surface of the two adsorbents. The values of Temkin's constant related to binding energy are high at the surface of ACMSA (0.9051/mg and 0.65991/mg) compared to those at the surface of AC (0.02681/mg and 0.00241/mg) which would reflect an increase in the binding strengths between the adsorbates and ACMSA [41]. That would be due to the increase in oxygen functions at the surface of the activated carbon following the functionalization with sulfuric acid. Values greater than 8KJ/mol of the binding energy provided by the D-R isotherm would indicate chemisorption [42]. However, the values of the Frumkin constant are all greater than zero, reflecting an attraction between the adsorbed species [43]; that attraction could be due to the presence in the aqueous medium of two different types of pollutants. However, even if the functionalization seems to have improved the adsorption at the surface of the modified activated carbon, the high value of the binding energy reflecting chemical adsorption and the fact that the Freundlich isotherm is the one that best describes the Adsorption of Cr(VI) to the surface of AC reveals an effect contrary to the expected effect.

## **I. Kinetic**

The pseudo-first-order, pseudo-second-order, Elovich, and intraparticle kinetic models were studied and the values of parameters provided by the latter are shown in table no 4.

**Table no. 4 : Kinetic parameters of Cr(VI) and orange methyl adsorption onto ACMSA and AC**

ACMSA		AC		
Cr	MO	Cr	MO	
<b>Pseudo-first-order</b>				
$Q_e(\text{mg/g})$	50.9069	288.5591	109.0000	-49.2691
$K_1$	0.0101	0.0645	-4.6294	
$R^2$	0.9362	0.6947	0.9336	
<b>Pseudo 2<sup>nd</sup> order</b>				
$Q_e(\text{mg/g})$	2.7405	1.7232	8.2919	2.7785
$K_2$	5.2216	-58.0600	0.6641	44.6648
$h(\text{mg/g/min})$	39.2157	-172.4100	45.6621	34.8275
$R^2$	0.9944	0.9830	0.5324	0.9782
<b>Elovich</b>				
$\beta(\text{g/mg})$	0.0104	0.0230	0.0489	0.0411
$\alpha(\text{mg/g/min})$	6.8398	11.2268	8.4117	8.4096
$R^2$	0.9888	0.9905	0.5665	0.9929
<b>Intra particular</b>				
$K_p(\text{mg/g/min}^{1/2})$	4.2448	1.7730	11.7780	12.5470
$C$	4.3518	-21.8660	-10.9490	-11.3430
$R^2$	0.9771	0.9345	0.6815	0.9579

The pseudo-second-order and Elovich models have the highest  $R^2$  values for adsorption of Cr(VI) and orange methyl to the surface of modified activated carbon (0.9944 and 0.9905, respectively) while the models pseudo-first-order and Elovich are those which have the highest correlation coefficients for the adsorption of Cr(VI) and orange methyl to the surface of AC (respectively 0.6947 and 0.9929). The pseudo-second-order adsorption therefore best describes the adsorption of Cr(VI) to the surface of the modified activated carbon; according to the literature, this corresponds more to the chemical adsorption process. This shows that the modification of the surface of the activated carbon would be at the origin of chemical functions favoring strong adsorption of Cr(VI). Elovich model better describes the adsorption of orange

methyl onto ACMSA and AC. The adsorption of Cr(VI) onto AC is best described by the pseudo-first-order model. The values of  $\alpha$  and  $\beta$  are 8.4096mg/g/min and 0.0411g/mg for the adsorption of orange methyl onto AC and 11.2268mg/g/min and 0.0230g/mg for the adsorption of the latter to the surface of ACMSA. The increase of the initial rate adsorption  $\alpha$  and the decrease of the initial desorption rate  $\beta$  indicates that the acid treatment would be at the origin of the increase of the adsorbent capacity to adsorb more and hardly achieve desorption. Similar results are obtained for the adsorption of orange methyl onto bamboo waste and bamboo waste modified with acid [44]. The existence of the constant C found with the intra-particle model shows that this model is not the step controlling the adsorption mechanism [45].

## CONCLUSION

In this study, the impact of the functionalization of activated carbon with sulfuric acid was evaluated for the affinity of the latter towards an inorganic pollutant such as chromium (VI) and a pollutant organic like orange methyl. The modification of activated carbon with sulfuric acid increase the quantity of some functional groups on the surface of this activated carbon, namely phenols, carboxylic acids and lactonic acids. This modification is responsible for the reduction of the iodine number (from 1422.40mg/g to 381mg/g). The adsorption of chromium (VI) and methyl orange to the surface of CA is best described by the isotherm of Freundlich ( $R^2 = 0.9037$ ) and Temkin ( $R^2 = 0.9869$ ) while the adsorption kinetics of these two pollutants to the surface of that adsorbent is better described by the pseudo-first-order model ( $R^2 = 0.6947$ ) and Elovich ( $R^2 = 0.9929$ ). Following the acid treatment, the adsorption of Cr (VI) and orange methyl to the surface of the activated carbon modified is now better described by the Langmuir isotherm ( $R^2 = 0.9943$  and  $0.9562$ ) and the kinetics of Adsorption is best described by the pseudo-second-order ( $R^2 = 0.9944$ ) and Elovich isotherm( $R^2 = 0.9905$ ).

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